



Dissolved organic carbon characteristics in surface ponds from contrasting wetland ecosystems: a case study in the Sanjiang Plain, Northeast China

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Abstract. Dissolved organic carbon (DOC) is a significant component of carbon and nutrient cycling in fluvial ecosystems. Natural wetlands, as important DOC sources for river and ocean ecosystems, have experienced extensive natural and anthropogenic disturbances such as climate change, hydrological variations and land use change in recent years. The DOC characteristics in surface ponds from contrasting wetlands in the Sanjiang Plain, Northeastern China were investigated. Surface ponds at seven sites (two natural phialiform wetlands, three natural riparian wetlands, one degraded wetland and one artificial wetland, i.e., rice paddy) were monitored during the growing seasons of 2009 and 2010. The results show that the surface ponds at the five natural wetland sites exhibited a wide range of DOC concentrations ($10.06\text{--}48.73\text{ mg L}^{-1}$) during the study period. The DOC concentrations showed no annual differences ($P > 0.05$) at all the wetland sites, except one of the phialiform wetland sites. The two phialiform wetlands exhibited higher DOC concentrations than the three riparian wetlands ($P < 0.05$). The DOC concentration in the surface pond at the artificial wetland site was relatively low ($P < 0.05$) compared to that at the degraded wetland site. The C/C ratios (the color per carbon unit ratio, $\text{Abs}^{400}/\text{DOC}$ concentration) showed inconsistent variations among these seven wetland sites, while the E4/E6 ratio ($\text{Abs}^{465}/\text{Abs}^{665}$, fulvic acid/humic acid) from the surface pond in the rice paddy land exerted 42.07–55.36% reductions ($P < 0.05$), compared to those at the five natural wetland sites. Furthermore, the E4/E6 ratio in the surface pond at the rice paddy site was significantly lower compared to that at the degraded wetland site ($P < 0.05$), which indicated that disturbance to wetland DOC in surface ponds might

be stronger when natural wetlands were converted to rice paddies in comparison with wetland degradation. This study could not only provide insightful points for understanding the aquatic DOC dynamics from different wetland ecosystems, but also support data information for incorporating the aquatic DOC into the model for regional carbon budgets in the future.

1 Introduction

Dissolved organic carbon (DOC) is a collective term for dissolved and colloidal organic compounds in various stages of decomposition, defined as the organic fractions that pass through a $0.45\text{ }\mu\text{m}$ filter (Roulet and Moore, 2006; Clay et al., 2009). It consists of a variety of molecules that range in size and structure from simple acids and sugars to complex humic substances (Thurman, 1985; Wallage et al., 2006). These different fractions of DOC differ in relation to the spectroscopic properties (Kalbitz et al., 2003; Hudson et al., 2007; Fellman et al., 2008). Previous studies have used a wide range of absorbances to measure DOC chemical characteristics (Thurman, 1985; Weishaar et al., 2003; Wallage et al., 2006; Fellman et al., 2008; Krupa et al., 2012). For example, specific UV (ultra-violet) absorbance (e.g., 280 nm) can be used for the initial assessment of the contribution of complex compounds to DOC (Rostan and Cellot, 1995; Kalbitz et al., 2003). The specific UV light absorption at 254 nm (SUVA_{254}) has been suggested as an index of the aromaticity of DOC (Weishaar et al., 2003; Fellman et al., 2008).

Fluorescence spectroscopy can provide an excellent tool to identify DOC fractions (Hudson et al., 2007).

These qualities of absorbance characteristics tend to vary with different sources of DOC from variable sites (Ågren et al., 2008). For instance, Höll et al. (2009) reported that specific UV absorbance values (280 nm, SUVA₂₈₀) of DOC showed significant difference in seasons or different sites (winter-rewetted site to winter moderately drained site, summer-rewetted site to summer moderately drained site). Natural wetland ecosystems store a substantial amount of carbon (Post et al., 1982; Mitra et al., 2005), despite only occupying 4 to 6 % of the Earth's land area (Matthews and Fung, 1987; Aselmann and Crutzen, 1989). Natural wetlands are important sources of DOC to the fluvial environment (Koprivnjak and Moore, 1992; Worrall and Burt, 2005; Wallage et al., 2006). DOC concentrations in streams draining 42 catchments showed consistent relationships with the variable percent wetland in the catchment (Eckhardt and Moore, 1990). Though wetland ecosystems have been reported to be highly correlated with river DOC export (Gergel et al., 1999; Freeman et al., 2001; Fellman et al., 2008), studies on the chemical quality of DOC from different wetland types are less (Fellman et al., 2008).

Wetland conversion activities in the Sanjiang Plain were extensive in the past 50 yr (Song et al., 2008). The conversion of natural ecosystems to artificial wetlands (i.e., rice paddy land) in this region has been accompanied by wetland degradation caused by the decrease of standing water depth and the input of nutrients during agricultural activities (Song et al., 2011) as well as the changing climate (Qian and Ruby Leung, 2007). Our previous study found that the DOC concentration from the surface pond decreased from the natural phialiform wetlands to rice paddy lands (Wang et al., 2010), while there was no significant difference between the riparian wetland and degraded wetland (Song et al., 2011). However, these preliminary findings are insufficient, thus, here a continuous two-year DOC monitoring was installed to examine whether the observed variations in DOC concentrations were maintained. Furthermore, we aimed to compare the DOC concentrations between (1) the natural riparian wetland and rice paddy land, (2) the DOC differences between phialiform wetlands and degraded wetlands, and lastly, (3) to compare DOC concentrations from rice paddy land and those from degraded wetland to clarify which exerts more influence on DOC dynamics. Especially, we want to figure out the DOC spectral characteristics among these different sites. Here absorbance at 400 nm was used, and also at 465 nm and 665 nm (Thurman, 1985) to represent the DOC color characteristics (Thurman, 1985; Wallage et al., 2006). A wide variety of studies have predicted DOC concentrations based on water color measurements (Moore and Jackson, 1989; Tao, 1998; Worrall et al., 2002, 2003; Worrall and Burt, 2005). This study aimed to better understand the aquatic DOC dynamics from different wetland ecosystems in the context of land use change and climate change.

2 Site description and methods

2.1 Site description

The Sanjiang Plain is framed by the Heilong River, Wusuli River and Songhua River (Fig. 1); moreover, many tributary rivers, such as Nongjiang River, Yalv River, Bielahonghe River and Naoli River exist in this area. The riparian wetlands developed from the flood plains near the rivers in this region. In addition, there are phialiform wetlands originating from the lowlands around the terrace, one of the typical geomorphic types (Liu and Ma, 2002). The geomorphic characteristics, boreal climate and hydrogeological conditions lead to the large areas with distributed natural wetlands in this region. Air temperatures in the Sanjiang Plain increase gradually from January to August, and then decline until December. Nearly half of the annual rainfall occurs in July and August. The Sanjiang Experimental Station of Wetland Ecology, Chinese Academy of Sciences, is located in the Sanjiang Plain. The climate in the station represents the regional weather pattern. The site-specific weather observations in the station showed that the mean air temperatures from June to September were 17.83 °C in 2009 and 19.79 °C in 2010. The total rainfall during this period was 509.40 mm for 2009 and 414.70 mm for 2010, respectively. This indicates that the general weather during our study period was warmer and drier in 2010 than that in 2009 (Fig. 2).

In this study we chose two typical phialiform natural wetlands – *Calamagrostis angustifolia* phialiform wetland (CAPW), and *Carex lasiocarpa* phialiform wetland (CLPW), three riparian natural wetlands near three tributary rivers – Bielahonghe riparian wetland (BRW), Nongjiang riparian wetland (NRW) and Yalv riparian wetland (YRW), the artificial wetland (AW: rice paddy land) and the degraded wetland (DW). The CAPW located in the outside position with a higher terrain is seasonally inundated, while the CLPW with lower terrain is permanently inundated with higher water level (Table 1). The hydrological conditions in the riparian wetlands are largely affected by rivers. The riparian wetlands show more drastic hydrological fluctuations. The contents of soil total nitrogen and soil organic matter (SOM) in the riparian wetlands are lower compared with the two phialiform wetlands. Compared with natural wetlands, the water level and SOM in the disturbed wetlands (DW and AW) decreased (Table 1). (NH₄)₂HPO₄ and urea are fertilized in the rice paddy land during the rice growth season. Hydrological and chemical characteristics of the study sites are listed in Table 1.

2.2 Water sample collection and DOC analysis

From April to September in 2010, samples from surface ponds were collected from CAPW, CLPW, BRW, NRW, YRW, DW and AW. Triplicate samples were collected in acid-washed 600 mL glass vials from these different types

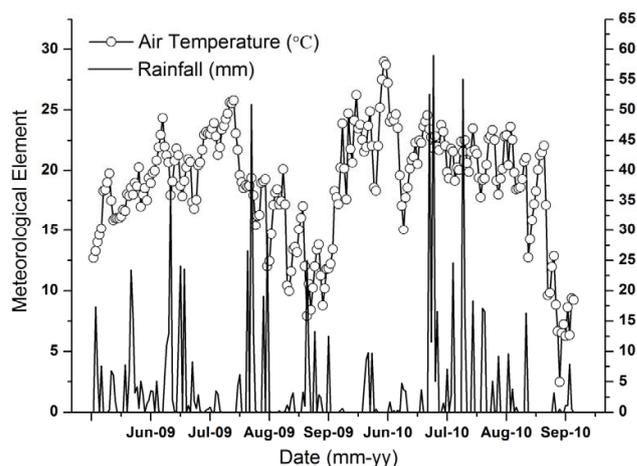
Table 1. Characteristics of the seven studying sites in the Sanjiang Plain, Northeastern China.

| Site | Water level (cm) | Total N (g kg ⁻¹) | SOM (%) | Soil total N (g kg ⁻¹) |
|------|------------------|-------------------------------|--------------|------------------------------------|
| CAPW | 12.49 ± 2.49 | 1.68 ± 0.27 | 9.47 ± 0.43 | 6.67 ± 0.61 |
| CLPW | 23.79 ± 1.71 | 1.29 ± 0.23 | 20.77 ± 0.71 | 21.17 ± 1.17 |
| BRW | 23.99 ± 4.51 | 0.69 ± 0.09 | 7.30 ± 0.23 | 5.77 ± 0.18 |
| NRW | 24.61 ± 1.48 | 1.01 ± 0.24 | 7.43 ± 0.44 | 5.23 ± 0.17 |
| YRW | 25.52 ± 1.94 | 0.74 ± 0.14 | 7.73 ± 0.22 | 5.83 ± 0.33 |
| DW | 8.83 ± 0.49 | 0.64 ± 0.24 | 5.57 ± 0.38 | 3.73 ± 0.19 |
| AW | 8.75 ± 0.79 | 1.17 ± 0.41 | 4.07 ± 0.23 | 2.67 ± 0.15 |

Note: *Calamagrostis angustifolia* phialiform wetland (CAPW), and *Carex lasiocarpa* phialiform wetland (CLPW), Bielahonghe riparian wetland (BRW), Nongjiang riparian wetland (NRW) and Yalv riparian wetland (YRW), the artificial wetland (AW-rice paddy land) and the degraded wetland (DW); the abbreviations in the following table source from the same full names.

**Fig. 1.** The location of the Sanjiang Plain, Northeast China (Wu, 2009).

of wetlands. Three sampling replicates at about 20 m intervals were collected from each site. For every sampling bottle, the sampled water consisted of water from four directions. In this study, the data of DOC concentrations for 2009 were cited from Wang et al. (2010) and Song et al. (2011). We collected topsoil (0–20 cm) with three soil profiles on each plot in September 2010 to analyze SOM and soil total nitrogen. The water samples were transported to the laboratory immediately after sampling, and were filtered through 0.45 µm filters. The first 100 mL of each sample was used to rinse the filter three times. The remaining was filtered into separate vials without any head space to minimize degassing (Wang et al., 2010; Song et al., 2011). Multi N/C 2100 Analyzer (Analytik Jena AG, Germany) was used to analyze total organic carbon (TOC) and total inorganic carbon (TIC) of the water samples. The analyzer used a non-dispersive infrared detector to quantitatively measure CO₂ levels to get the concentrations of TIC and TOC. DOC concentrations were determined by TOC minus TIC. Each sample was injected at least two times to obtain a standard deviation of ≤ 2%.

**Fig. 2.** Air temperature and rainfall in the Sanjiang Plain Wetland Station during the growing seasons of 2009 and 2010.

2.3 Determination of specific absorbance values and data analysis

The C/C ratio (the color per carbon unit ratio, Abs⁴⁰⁰/DOC concentration) of the samples was obtained by dividing the absorbance values at 400 nm (Abs⁴⁰⁰) by the corresponding DOC concentrations (Wallage et al., 2006). A high C/C ratio indicates that the DOC comprises a greater proportion of colored humic substances compared to uncolored non-humic substances (Wallage et al., 2006). The E4/E6 ratio was determined by dividing the absorbance at 465 nm (Abs⁴⁶⁵) by that at 665 nm (Abs⁶⁶⁵) for the individual samples (Wallage et al., 2006). This ratio can be used to measure the proportion of fulvic acid to humic acid in the colored humic component of DOC (Thurman, 1985; Wallage et al., 2006). The absorbance measurements of 400 nm, 465 nm and 665 nm were made on a UV-7504 spectrophotometer using deionized water as a blank. A quartz cell with a 1.0-cm path length was utilized.

In addition, the SPSS 11.5 software and Origin 7.5 statistical packages were used in the statistical analysis. The difference in DOC, C/C and E4/E6 ratios among different sites was tested by the repeated measures analysis of variance tests (ANOVAs). In analysis where $P < 0.05$, the comparisons were considered statistically significant.

3 Results

3.1 DOC concentrations

During the study period, both inter-seasonal and extreme inter-annual variations of DOC concentrations were identified in the surface ponds in different types of wetlands (Fig. 3). The averaged values for DOC during the two-year samplings ranged from 7.08 ± 1.38 to 48.73 ± 3.26 mg L⁻¹

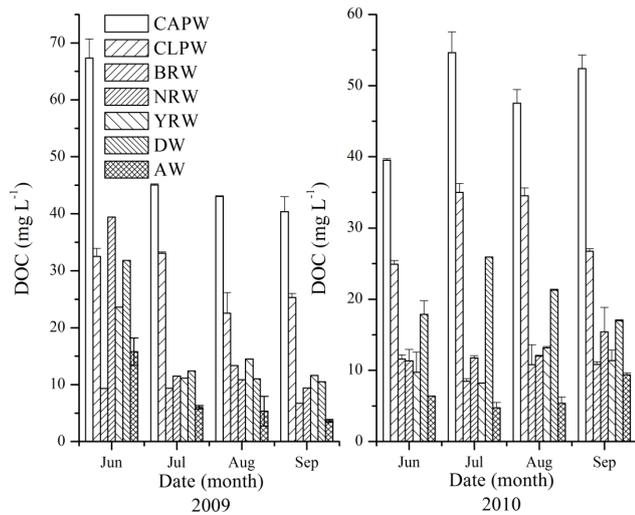


Fig. 3. Monthly DOC concentrations in surface ponds from contrasting wetland ecosystems in the Sanjiang Plain during the growing seasons of 2009 and 2010. Note: the data for 2009 is cited from Wang et al. (2010) and Song et al. (2011); *Calamagrostis angustifolia* phialiform wetland (CAPW), and *Carex lasiocarpa* phialiform wetland (CLPW), Biehanghe riparian wetland (BRW), Nongjiang riparian wetland (NRW) and Yalv riparian wetland (YRW), the artificial wetland (AW-rice paddy land) and the degraded wetland (DW); the abbreviations in the following figures source from the same full names.

at seven sites. The highest two-year mean DOC concentration was found in the surface pond at the CAPW site, followed by the CLPW site ($29.33 \pm 1.75 \text{ mg L}^{-1}$), the DW site ($18.48 \pm 2.67 \text{ mg L}^{-1}$), the NRW site ($15.21 \pm 3.51 \text{ mg L}^{-1}$), the YRW site ($12.91 \pm 1.67 \text{ mg L}^{-1}$) and the BRW site ($10.06 \pm 0.72 \text{ mg L}^{-1}$), while the lowest value was observed in the surface pond at the AW site. The two-year DOC concentrations in the surface pond from CAPW were higher than that in CLPW ($P < 0.05$). There was no significant difference among the three riparian natural wetlands ($P > 0.05$). The two-year DOC concentrations in surface ponds from both CAPW and CLPW sites were significantly higher compared to those in the three riparian natural wetlands, the degraded wetland and the rice paddy land ($P < 0.05$). The DOC concentrations at the BRW site was higher than those at the degraded wetland site ($P < 0.05$), yet there were no significant differences between the other two riparian wetlands (NRW and YRW) and the DW. The DOC concentrations at the NRW site were significantly higher than that at the AW site, while no great differences were observed between the other two riparian wetlands (BRW and YRW) and the AW ($P > 0.05$). It is notable that the DOC concentrations at the DW site were remarkably higher than that at the AW site ($P < 0.05$). As for the between-year difference, the DOC concentration in the surface pond at the CLPW site in 2009 was greater than that in 2010 ($P < 0.05$). No significant differences in DOC concentrations were observed between

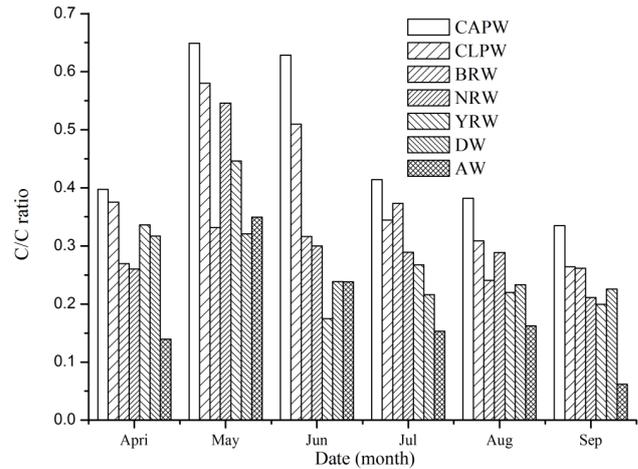


Fig. 4. Monthly C/C ratio in surface ponds from contrasting wetland ecosystems in the Sanjiang Plain during the growing season of 2010.

2009 and 2010 for the surface ponds at the other six sites ($P > 0.5$).

3.2 C/C ratio

The C/C ratios ranged from 0.06 to 0.65 in the surface ponds we sampled from April to September in 2010 (Table 2 and Fig. 4). The C/C value at the CAPW and CLPW sites showed little difference ($P > 0.5$). Also, no significant C/C difference between the three riparian wetlands was observed ($P > 0.5$). The C/C value at the CAPW site is greater than those at the three riparian wetland sites, DW and AW sites ($P < 0.5$), while the C/C at the CLPW site is much higher than those at the YRW, DW and AW sites. The C/C value in the degraded wetland was lower compared to those in the three riparian wetlands ($P < 0.5$), and the C/C ratios were almost the same in the three riparian sites ($P > 0.5$). The ratio at the NRW site was relatively high compared to that at the AW site, but we observed no large differences between the other two riparian sites and the AW sites ($P > 0.5$). In addition, the differences between the DW and AW sites can be ignored ($P > 0.5$).

3.3 E4/E6 ratio

The E4/E6 ratio of DOC in surface ponds under different types of wetlands is presented in Table 2 and Fig. 5. The E4/E6 ratio in the surface ponds of the five natural wetlands showed no significant differences ($P > 0.5$). Meanwhile, this ratio in the surface pond at the DW site was almost the same as those at the natural wetland sites ($P > 0.5$), but the seasonal averaged E4/E6 ratio in the surface pond at the AW site was lower than those at the natural wetland sites ($P < 0.5$). The E4/E6 ratio in the surface pond at the AW site showed 42.07–55.36 % reductions ($P < 0.5$) compared to those at the

Table 2. Average values of C/C and E4/E6 ratios in surface ponds from contrasting wetlands in the Sanjiang Plain during the growing season of 2010.

| | | Sites | | | | | | |
|-------------|---------|--------------|---------------|----------------|---------------|---------------|---------------|--------------|
| | | CAPW | CLPW | BLRW | NLRW | YLRW | DW | AW |
| C/C ratio | Range | 0.33–0.65 | 0.26–0.58 | 0.24–0.37 | 0.21–0.55 | 0.17–0.45 | 0.22–0.32 | 0.06–0.35 |
| | Average | 0.47 ± 0.06a | 0.40 ± 0.05ab | 0.30 ± 0.02bcd | 0.32 ± 0.05bc | 0.27 ± 0.04cd | 0.26 ± 0.02cd | 0.18 ± 0.04d |
| E4/E6 ratio | Range | 4.27–6.45 | 6.07–7.00 | 4.67–11.67 | 4.00–11.00 | 4.19–7.5 | 5.25–9.00 | 2.33–3.62 |
| | Average | 5.71 ± 0.32a | 6.53 ± 0.16a | 7.28 ± 1.16a | 6.98 ± 0.95a | 5.61 ± 0.59a | 6.21 ± 0.58a | 3.25 ± 0.21b |

Note: Average is (mean ± SE); there is significant difference among those averaged values with different lower-alpha after the averaged values, whereas no significant difference.

natural wetland sites. Furthermore, the E4/E6 ratio in the surface pond at the AW site revealed relatively low values compared to that at the DW site ($P < 0.5$).

4 Discussion

4.1 Comparisons with other studies

Natural wetlands, as significant carbon pools, have been increasingly recognized as important DOC sources for adjacent waters (Wallage et al., 2006; Dawson and Smith, 2007; Baker et al., 2008). Much work has been completed concerning the concentration and bioavailability of soil dissolved organic carbon from wetland ecosystems in the Sanjiang Plain (Zhang et al., 2005; Zhang et al., 2008; Guo et al., 2010), yet rare information is available about how the dynamics of DOC are derived from surface ponds in wetland ecosystems of the Sanjiang Plain (Wang et al., 2010; Song et al., 2011). In this study we determined that the averaged DOC value was 23.25 mg L^{-1} for the surface ponds at the five natural wetland sites in the Sanjiang Plain during the two-year samplings. Nagao et al. (2007) showed that DOC concentrations in river waters from tributaries of the Amur River in the Sanjiang Plain were generally lower than 10 mg L^{-1} . The higher concentrations we recorded suggest that the natural wetlands in our study could serve as alternative DOC sources for tributaries along the Amur River in the Sanjiang Plain. Similarly, Waiser and Robarts (2004) reported that natural wetlands across the Canadian prairies exhibited high DOC concentrations ($> 10 \text{ mg L}^{-1}$), acting as significant DOC pools for adjacent waters. Raphael et al. (1996) and Arrigoni et al. (2008) pointed out that the Tivoli Bay's freshwater tidal wetlands exported DOC to the main channel of the Hudson River, but the magnitude of DOC concentrations ($< 7 \text{ mg L}^{-1}$) in their studies were much lower than in our study. These variations might result from different rates of DOC production, solubility and transport under freshwater and tidal wetlands. Alternatively, this study together with previous studies provided substantial evidence for the importance of natural wetlands on DOC export to adjacent rivers. This is also confirmed by Hope et al. (1997) who reported a significant correlation between the average of DOC concentration in 11 Scottish catchments and the percentage of natural wetland cover.

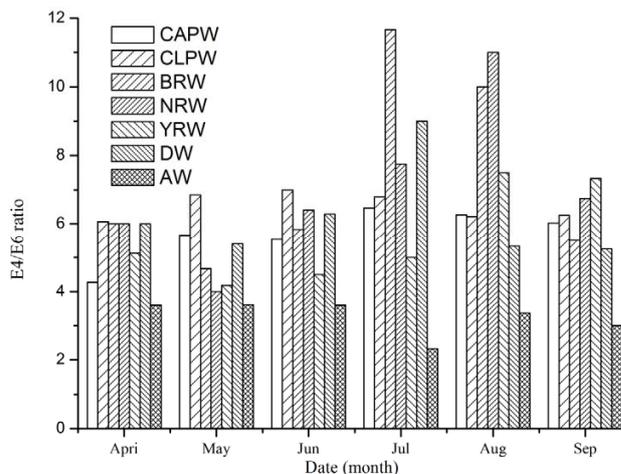


Fig. 5. Monthly E4/E6 ratios in surface ponds from contrasting wetland ecosystems in the Sanjiang Plain during the growing season of 2010.

4.2 DOC characteristics from different natural wetland ecosystems

In this study, we found that DOC concentrations in the surface ponds at the natural sites showed the decreasing sequence of CAPW > CLPW > riparian wetlands during our two-year study ($P < 0.5$). The greater DOC values from the surface ponds of phialiform wetlands than those in the riparian wetlands might result from the varying hydrological conditions in the two wetland types. Compared with the two phialiform wetlands, riparian wetlands were flooded with higher water levels (Table 1), which would lead to more anaerobic conditions, lower soil carbon decomposition rates, and thus less release of DOC from soil and litter decomposition in the surface ponds. Another potential driver of the observed differences might be the contrasting nutrient contents of the wetlands. DOC is a product of SOM, and other nutrients (i.e., soil total nitrogen, soil total phosphorus, etc.) could control the DOC production/consumption processes (Chapin et al., 2002). Although no correlations between nutrient factors and DOC concentrations were investigated in this study, the total nitrogen in water, soil total nitrogen and SOM showed higher values at the phialiform wetland sites (Table 1) than those at the riparian wetland sites. Studies in

UK have also reported that the DOC releases from different wetland types depended on nutritional grounds (Mitsch and Gosselink, 1993; Freeman et al., 2004). The significant differences in two-year DOC concentrations between CAPW and CLPW, which is also confirmed by Wang et al. (2010), might come from the fact that they held different contents of soil organic carbon (Table 1; Zhang, 2006) along with different hydrologic characteristics (Wang et al., 2010). Furthermore, soil pH values from wetland ecosystems showed a contrary trend compared to soil organic carbon (SOC) changes (Yang, 2006), and also affected the microbial species (Li et al., 2001), which might influence DOC variations from surface ponds in wetland ecosystems.

In addition to DOC concentrations, this study investigated color/carbon ratios (C/C). The C/C ratios showed no significant difference between the two phialiform wetland sites, while the C/C ratios were higher at the CAPW sites compared with those at the three riparian wetland sites. This suggests that, for every carbon unit (e.g., per mg C), the DOC at the CAPW site contained higher proportions of colored humic substances compared to uncolored non-humic substances (Wallage et al., 2006) than per unit of carbon at the riparian sites. However, the E4/E6 differences between the two phialiform wetland sites and the riparian wetland sites were not significant. It implied that the fulvic acid/humic acid ratios of these two wetland types are similar.

4.3 DOC dynamics from artificial wetland and degraded wetland

In the present study, DOC concentration in the surface pond at the artificial wetland is much lower than the degraded wetland site ($P < 0.5$). It implied that the DOC concentrations leaching to water bodies was lower from rice paddy land compared with degraded wetland. This might be because the agricultural management practices (e.g., fertilizer use and irrigation) within the rice paddy land reduce DOC concentrations in water bodies. It has been reported that additions of acid anions like sulphate (SO_4^{2-}) and nitrate (NO_3^-) to wetland soils could lead to DOC consumption during redox reactions (Bartlett et al., 2005). In addition, compared with the two phialiform wetlands, we found DOC concentration depletion in the rice paddy land, which is also verified by our precious study (Wang et al., 2010). Special attention of DOC concentrations should be paid on the degraded wetlands, which showed lower DOC values than those of the two phialiform counterparts. However, the DOC concentrations from surface ponds in the rice paddy land or degraded wetland only showed significant differences with one of the three riparian wetlands. This difference shows that highly variable ecosystem functions in different wetland types (Groffman et al., 1996) might be taken into account, with regard to DOC changes disturbed by human activities.

The E4/E6 values during our study period were the lowest in the surface pond at the artificial site compared to those at

the phialiform wetland sites, the riparian wetland sites, and the degraded wetland sites ($P < 0.5$). This clear difference in the E4/E6 ratios between the artificial wetland site and the other sites implied that DOC at the rice paddy site was less dominated by immature fulvic acids compared with those at the other sites. It might be caused by the relatively low level of microbial activity (Wallage et al., 2006) in the rice paddy land. It has been reported that the microbial biomass carbon significantly decreased after natural wetlands were converted to rice paddy lands (Zhang, 2006). There are no significant differences in the E4/E6 ratio between the surface pond at the natural wetland sites and the degraded sites. This indicates that natural wetland conversion to rice paddies has stronger effects on aquatic DOC chemical characteristics in comparison with wetland degradation.

5 Conclusions

In this study, surface ponds from the five natural wetlands in the Sanjiang Plain showed wide ranges of DOC concentrations (10.06–48.73 mg L⁻¹), C/C ratios (0.27–0.47) and E4/E6 ratios (5.61–7.28). Different scenarios of DOC variations among the two phialiform wetlands and the three riparian wetlands were introduced, compared with artificial wetland or degraded wetland. In the various wetland ecosystems in northeast China, ignoring the wetland types would lead to biases of the DOC loss potential with land use change and other environmental stresses. In addition, lower DOC concentrations in the surface pond of the artificial wetland than that in the degraded wetland were observed. Lower E4/E6 ratio was also found in the artificial wetland than those at the other six sites, though no differences were observed between natural sites and degraded sites. Natural wetland conversion to rice paddies might exert stronger effects on aquatic DOC chemical characteristics compared with those from wetland degradation. Additional site-specific field measurements are required to further refine the differences in DOC characteristics among the natural sites, arable land sites and degraded sites in Northeastern China, and the corresponding parameters, such as SOC, pH-values or redox conditions need to be monitored to identify the underlying reasons for DOC dynamics.

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